

VARIATION IN THE HERBACEOUS-SUBSHRUB STRATUM AS AN EFFECT OF FRAGMENTATION IN THE CERRADÃO

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ABSTRACT

Habitat fragmentation favors the invasion of exotic species. Cerradão is the most fragmented phytophysognomy in the Brazilian savanna. Thus, to provide insights into the effects of fragmentation in the cerradão, we compared the composition and structure of the herbaceous-subshrub layer on the edge and inside a fragment of cerradão located in the Federal District, Brazil. Vegetation coverage, including floristic richness and diversity, was higher at the edge, where all registered exotic species were present. Invasive *Melinis minutiflora* and *Urochloa brizantha* were the most dominant species on the edge of the cerradão, while the native *Echinolaena inflexa* and *Rhynchospora exaltata* were dominant in the interior. Invasive species accounted for more than 10% of the total species observed at the edge and contributed to more than 50% of their live cover. Therefore, the studied cerradão seems to suffer from the effects of fragmentation. Further studies in the area are required.

Keywords: Coverage, Edge, Frequency, Invasive species, Savanna.

RESUMO

A fragmentação do habitat favorece a invasão de espécies exóticas. Cerradão é a fitofisionomia mais fragmentada da savana brasileira. Assim, para fornecer informações sobre os efeitos da fragmentação no cerradão, comparamos a composição e a estrutura da camada herbácea-subarbusciva na borda e no interior de um fragmento de cerradão localizado no Distrito Federal, Brasil. A cobertura vegetal, assim como a riqueza e diversidade florística, foi maior na borda, onde todas as espécies exóticas registradas estavam presentes. As invasoras *Melinis minutiflora* e *Urochloa brizantha* foram as espécies mais dominantes na borda do cerradão, enquanto as nativas *Echinolaena inflexa* e *Rhynchospora exaltata* foram dominantes no interior. As espécies invasoras representaram mais de 10% do total de espécies observadas na borda e contribuíram com mais de 50% de sua cobertura viva. Com isso, o cerradão estudado parece sofrer os efeitos da fragmentação. No entanto, novos estudos na área são necessários para confirmar tais efeitos.

Palavras-chave: Cobertura, Borda, Frequência, Espécies invasoras, Savana.

INTRODUCTION

Habitat fragmentation is the reduction or loss of area and its division into smaller and isolated parts. The smaller the fragment, the greater the amount of edge and the

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vulnerability to anthropogenic pressures (Primack & Rodrigues, 2001). Fragmentation can result from anthropogenic or natural activities, and both have very different ecological consequences. Anthropogenic fragments tend to possess low biodiversity and pose a threat to the population viability of species, mainly because of the scarcity of essential natural resources. Conversely, natural fragments are more likely to contain endemic species because of their longer isolation time (Constantino *et al.*, 2005; Aquino & Miranda, 2008).

Native vegetation refuges in man-altered spaces behave like islands for species that are dependent on natural habitats (Diamond, 1976). An island's degree of isolation and ecological dynamics affect the richness of local species. Large refuges are essential to minimize extinction rates, since small fragments tend to lose species that are dependent on well-established habitats, and retain successional and border species, which are characterized by rapid dispersion (MacArthur & Wilson, 2016). However, it is important to consider the spatiotemporal scale and the vegetation type analyzed, as landscapes are formed by mosaics due to the determination of ecosystems by edaphic, climatic, and historical conditions (Aquino & Miranda, 2008).

The Brazilian savanna is the second largest biome in Brazil, and is characterized by a vegetation complex that has ecological and physiognomic relationships with other savannas in tropical America, Africa, Southeast Asia, and Australia. Its extensive distribution in both latitude and altitude results in wide thermal diversification (Ribeiro & Walter, 2008). The Brazilian savanna is a mosaic of physiognomic forms (Coutinho, 2006), and edaphic, geomorphological, topographic, and climatic factors are particularly significant.

Although few long-term studies have investigated the ecological processes lost due to anthropic fragmentation in the Brazilian savanna, the main responses of this biome to fragmentation include an increase in species richness, principally exotic species, and the rapid recruitment of plants adapted to ambient disorders. These factors, together with the presence of ruderal native species, reflect the vulnerability of the fragment and indicate that its border is an altered environment (Aquino & Miranda, 2008; Dodonov *et al.*, 2016; Pereira *et al.*, 2021). Newly formed forest edges seem to favor the regeneration of seedlings due to the high light availability and seed deposition. Conversely, changes to the local microclimate result in the loss of native plant species, which are adapted to temperature, humidity, and luminosity, in addition to changes in the distribution of insects and in plant–insect interactions, specifically in Brazilian savanna (Christianini & Oliveira, 2013; Dodonov *et al.*, 2013; Mendonça *et al.*, 2015).

With more than 4,800 endemic species of plants and vertebrates, the Brazilian savanna is a global biodiversity hotspot (Strassburg *et al.*, 2017). However, despite its importance for the conservation of species and the provision of ecosystem services, to date, the Brazilian savanna has lost almost 50% of its native vegetation cover, with only 19.8% remaining intact and less than 8% protected in public conservation units (Beuchle *et al.*, 2015), mainly because of the strong pressure for agricultural expansion (Strassburg *et al.*, 2017).

Among the various phytophysionomies in the Brazilian savanna, which encompass forestry, savanic, and field formations, the cerrado is the most fragmented and degraded formation due to human activities, mainly due to its location on flat topography, fertile, and arable land. The cerrado is a woody savanna that occurs in the interfluvia in well-drained land, without association with water courses (Ribeiro and Walter, 2008; Miguel *et al.*, 2017). According to Ribeiro & Walter (2008), cerrado is a forest phytophysionomy with xeromorphic aspects, and is characterized by the presence of species that occur both in the savanna and in the dry forest. It varies between 8 and 15 m in height, and has a predominantly continuous canopy and tree cover that varies from 50 to 90%, providing light conditions that favor the formation of differentiated shrub and herbaceous strata.

Due to the size of the forest and the predominance of tree species, the areas they occupy have been used frequently for agricultural and carpentry purposes, resulting in the occurrence of fragments that are generally located between areas dominated by the savanna (Sano *et al.*, 2010). Studies focused on the phytosociology of the herbaceous-subshrub stratum in forest environments are rare because the arboreal component holds the most biomass and has great economic importance, although smaller species are highly sensitive to microclimatic and edaphic variations and are relevant for ecology and conservation studies (Meira-Neto & Martins, 2000; Miyamura *et al.*, 2019).

The more fragmented and altered a landscape, the greater the chances of introducing alien species (Mendonça *et al.*, 2015). For example, in the Brazilian savanna, exotic grasses pose a risk to biodiversity because they prevent the establishment of native grasses; however, plant phenology is also influenced by changes in abiotic conditions caused by habitat fragmentation, such as increased luminosity and temperature (Vogado *et al.*, 2016). In a recent study, Tallamy *et al.* (2020) reported that the expansion of non-native plant species is responsible for the decline in populations of native herbivorous insects. This contributes to the trophic imbalance of the biome, as these species are important components of the diet for most terrestrial bird species, which are important dispersers in the Brazilian savanna.

Thus, to provide insights into the effects of fragmentation on cerrado phytophysiognomy, this study compared the composition and structure of the herbaceous-subshrub layer at the edge and inside of a cerrado fragment located in the Federal District, Brazil. The vegetation in this stratum is expected to be more dense at the boundary of the fragment due to colonization by exotic and invasive species.

MATERIAL AND METHODS

Study area

The Cerradão Biological Reserve (“REBIO do Cerradão”) (Figure 1) is an Integral Protection Conservation Unit that was created as an Area of Relevant Ecological Interest by decree no. 19.213/98 (Distrito Federal, 1998). It was transformed into a Biological Reserve through decree nº 31.757/10 (Distrito Federal, 2010) because of its environmental and ecological characteristics, which are important for the conservation of phytophysiognomy. According to the SNUC (Brasil, 2000), the objective of a Biological Reserve is the integral preservation of biota and other natural attributes existing in its limits, without direct human interference or environmental modifications. Conversely, an Area of Relevant Ecological Interest is a small unit of sustainable use with little or no human occupation, which presents notable natural characteristics or houses rare specimens of the regional biota. This type of conservation unit allows the regulated use of its areas, as long as this use is compatible with the objectives of nature conservation.

The “REBIO do Cerradão” is located within the Environmental Protection Area (EPA) Gama and Cabeça de Veado (a conservation unit of sustainable use) at the coordinates 15°51'S and 47°49'W. To the east, it borders the Contorno Park Road (EPCT; DF-001), to the west, the Dom Bosco Mansions Sector (SMDB, 4th stage), to the north, the Adventist Hospital, and to the south, the Cabeça de Veado Park Road (EPCV; DF-035). It has a reservoir supported by the CAESB (Federal District Environmental Sanitation Company). The conservation unit occupies an area of 54.12 ha, including 16 ha of cerrado and adjacent dense savanna (i.e., dense cerrado *sensu stricto*). Its location on the border between the São Bartolomeu and Paranoá River Basins, the “REBIO do Cerradão” is important for the infiltration of water in the soil, and is present within a region that has potential for aquifer recharge of approximately 50% (Instituto Brasília Ambiental, 2014).

According to physico-chemical analyses performed in the cerradão, the soil is of the Red Latosol dystrophic type, with base saturation of 6%, average CTC (8.3 cmolc/dm^3) and 81.4 g/kg of organic matter. The soil texture of the cerradão is clayey (60%) and the values of Al, Ca, Mg, K, and Na are, 1.4, 0.3, 0.1, 0.13, and 0.01 cmolc/dm^3 , respectively. At pH 4.9, the cerradão is classified as strongly acidic, which characterizes dystrophic soils.

The local climate is tropical Aw according to the Köppen and Geiger classification, which is typical of the Federal District region, with two well-defined seasons: dry winter and rainy summer. The average annual temperature in the region is 21.1°C and the average annual rainfall is $1,668 \text{ mm}$ (Cardoso *et al.*, 2014; Climate-Data, 2020).

The CAESB Supported Reservoir was built and activated in mid-May 1978 (H.O. Antunes, personal communication). The fragments revealed no traces of a recent fire.

Sampling

The floristic composition of “REBIO do Cerradão” was determined through biweekly collections carried out from September 2007 to November 2008. The collected individuals were herborized and incorporated in the Herbarium of the University of Brasília (UB). The identification was revised and followed the botanical nomenclature proposed by the Angiosperm Phylogeny Group IV (2016). The synonyms were updated based on Flora do Brasil 2020 (Brazil Flora Group, 2021).

Phytosociological sampling of the herbaceous-subshrub stratum occurred at the beginning of the rainy season (November 2008). Two environments were determined along the cerradão: edge (the shortest distance from the fragment's limit) and interior (the longest distance from the fragment's limit). The 400 m external strip to the south of the fragment was considered as the edge environment. For the interior environment, the innermost area of the cerradão was considered, 175 m away from its limits, forming a sample area of $2,500 \text{ m}^2$ ($50 \times 50 \text{ m}$). The two environments were divided into $1 \times 1 \text{ m}$ squares. Each grid comprised one numbered sample unit (plot), and 40 numbers were randomly drawn in each of the areas using the Research Randomizer software (Urbaniak & Plous, 2007) (Figure 2). Only live biomass was considered in phytosociological sampling.

Coverage of the herbaceous-subshrub layer was estimated using the Braun-Blanquet scale (Kent, 2012), distinguishing the values of live and dead biomass in relation to the area of the plots and for each individual species. Herbaceous strata were considered herbs and non-woody lianas, and subshrub strata were considered the subshrub species and acaule palm trees. The species were classified as “naturalized exotic” following the information published by the Flora do Brasil 2020 (Brazil Flora Group, 2021).

Absolute frequency (FA) and relative dominance (DoR) were calculated using data obtained in the field and are presented as percentages. The occurrence of each species in the plots was used to calculate the frequency, and coverage was used to represent dominance. The formulas used for phytosociological calculations were adapted from Müller-Dombois and Ellenberg (2002), as follows:

a) Absolute Frequency: $FA = \frac{p_i}{\sum p} \cdot 100$

p_i = number of plots in which a given species occurred

p = number of plots sampled

b) Relative Dominance: $DoR = \frac{(c/ha)}{(C/ha)} \cdot 100$

c = coverage value of a given species

C = total area coverage value

where: $c = \sum c_{pi}$ (c_{pi} = coverage of the plot in which a given species occurred)

$C = \sum cp$ (cp = coverage of each plot)

Data analysis

The species diversity of the whole cerrado and each sampled environment (border and interior) was measured using the Shannon-Wiener index (H') (Kent, 2012), which is based on the proportional abundance of species, and assigns greater values to rare species. Equability in the distribution of species area was calculated using the Pielou uniformity index (J') (Pielou, 1966). This index is derived from the Shannon diversity index and represents the uniformity of the distribution of individuals among the existing species.

Similarity between the edge and interior plots was assessed using the Sørensen index (S_s) (Kent, 2012), which qualitatively compares two samples within the same community, assigning greater weight to species that are common to the areas being compared. The statistical program MVSP was used to calculate the three indexes (Kovach, 2007).

Finally, to compare the percentage coverage of the herbaceous-subshrub layer in the edge and interior environments, we used the non-parametric Kolmogorov-Smirnov (K-S) test (Zar, 2010), with a significance level (α) of 0.05. The PAST statistical program was used for the K-S test (Hammer *et al.*, 2008).

RESULTS

Vegetation richness and diversity

The floristic composition of the cerrado includes approximately 95% native species of the biome, 60% of families represented by a single genus, and 50.66% represented by a single species. These values, together with the Shannon ($H' = 3.85$ nats/ind.) and Pielou ($J' = 0.82$) indices reflect the high richness and diversity of species in the area (Silva & Felfili, 2012).

A survey of herbaceous-subshrub flora identified 109 species, including 50 herbs, 37 subshrubs, and 22 climbers belonging to 90 genera and 39 botanical families. Of these, 56.41% were represented by only one species. The best-represented families were Poaceae (16 spp.), Asteraceae (14 spp.), Rubiaceae (11 spp.), and Fabaceae (10 spp.). Of the total species identified, 6.42% were naturalized exotic (Table 1).

Table 1. Herbaceous-subshrub flora of the cerrado in the “REBIO do Cerradão”, Federal District, Brazil. H = herb. C = climber. S = subshrub. *: Exotic naturalized.

Family	Species	Habit	Voucher
Acanthaceae	<i>Justicia lanstyakii</i> Rizzini	H	UB109112
Alstroemeriaceae	<i>Alstroemeria gardneri</i> Baker	H	UB109465
Amaranthaceae	<i>Gomphrena agrestis</i> Mart	H	UB109578
	<i>Pfaffia denudata</i> (Moq.) Kuntze	H	UB109314
Anemiaceae	<i>Anemia phyllitidis</i> (L.) Sw.	H	UB109365
Apocynaceae	<i>Ditassa lenheirensis</i> Silveira	C	UB111984
	<i>Ditassa obcordata</i> Mart	C	UB109650
	<i>Odontadenia lutea</i> (Vell.) Markgr.	C	UB109473
	<i>Prestonia coalita</i> (Vell.) Woodson	C	UB109454
Arecaceae	<i>Allagoptera campestris</i> (Mart.) Kuntze	S	UB109299
Asteraceae	<i>Achyrocline satuireioides</i> (Lam.) DC.	S	UB109643
	<i>Aspilia foliacea</i> (Spreng.) Baker	H	UB109339
	<i>Baccharis dracunculifolia</i> DC.	S	UB109359
	<i>Campuloclinium megacephalum</i> (Mart. ex Baker) R.M.King & H.Rob.	S	UB109480
	<i>Echinocoryne holosericea</i> (Mart.) H.Rob.	H	UB109644

	<i>Elephantopus mollis</i> Kunth	H	UB109623
	<i>Emilia fosbergii</i> Nicolson *	H	UB109354
	<i>Hoehnephytum trixoides</i> (Gardner) Cabrera	S	UB109677
	<i>Ichthyothere latifolia</i> (Benth.) Gardner	S	UB109464
	<i>Lessingianthus compactiflorus</i> (Mart.ex Baker) H.Rob.	S	UB109458
	<i>Symphypappus compressus</i> (Gardner) B.L.Rob.	S	UB109388
	<i>Trichogonia salviifolia</i> Gardner	H	UB109355
	<i>Tridax procumbens</i> L. *	H	UB109482
	<i>Wedelia bishopii</i> H.Rob.	H	UB109507
Balanophoraceae	<i>Langsdorffia hypogaea</i> Mart.	H	UB109613
Bignoniaceae	<i>Anemopaegma acutifolium</i> DC.	S	UB109615
	<i>Cuspidaria sceptrum</i> (Cham.) L.G.Lohmann	C	UB109272
	<i>Fridericia platyphylla</i> (Cham.) L.G.Lohmann	C	UB109477
	<i>Jacaranda ulei</i> Bureau & K.Schum.	S	UB109199
Blechnaceae	<i>Blechnum auriculatum</i> Cav.	H	UB109370
Boraginaceae	<i>Varronia truncata</i> (Fresen.) Borhidi	S	UB109485
Bromeliaceae	<i>Dyckia</i> sp.	H	UB109616
Convolvulaceae	<i>Distimake contorquens</i> (Choisy) A.R.Simões & Staples	C	UB109563
	<i>Ipomoea procumbens</i> Mart. ex Choisy	C	UB109424
Cucurbitaceae	<i>Cayaponia tayuya</i> (Vell.) Cogn.	C	UB109439
Cyperaceae	<i>Cyperus laxus</i> Lam.	H	UB109366
	<i>Rhynchospora consanguinea</i> (Kunth) Boeckeler	H	UB109443
	<i>Rhynchospora exaltata</i> Kunth	H	UB109143
Dioscoreaceae	<i>Dioscorea dodecaneura</i> Vell	C	UB110707
Euphorbiaceae	<i>Dalechampia caperonioides</i> Baill.	H	UB109438
Fabaceae	<i>Bauhinia dumosa</i> Benth.	S	UB109675
	<i>Bauhinia rufa</i> (Bong.) Steud.	S	UB110711
	<i>Calliandra dysantha</i> Benth.	S	UB109139
	<i>Centrosema brasilianum</i> (L.) Benth.	C	UB109569
	<i>Cerradicola boavista</i> (Vell.) L.P.Queiroz	H	UB109310
	<i>Cerradicola grewiifolia</i> (Benth.) L.P.Queiroz	S	UB109468
	<i>Chamaecrista conferta</i> (Benth.) H.S.Irwin & Barneby	S	UB109181
	<i>Crotalaria flavicoma</i> Benth.	H	UB109472
	<i>Mimosa somnians</i> Humb. & Bonpl. ex Willd.	S	UB109441
	<i>Stylosanthes guianensis</i> (Aubl.) Sw.	H	UB109210
Gentianaceae	<i>Voyria aphylla</i> (Jacq.) Pers.	H	UB109495
Lamiaceae	<i>Hyptis brachystachys</i> (Pohl ex Benth.) Harley	S	UB109634
	<i>Hyptis rubiginosa</i> Benth.	S	UB109203
	<i>Hyptis villosa</i> Pohl ex Benth.	H	UB203032
	<i>Oocephalus lythroides</i> (Pohl ex Benth.) Harley & J.F.B.Pastore	S	UB109223
Lauraceae	<i>Cassytha filiformis</i> L.	C	UB109460
Loranthaceae	<i>Passovia ovata</i> (Pohl ex DC.) Tiegh.	H	UB109247
	<i>Psittacanthus robustus</i> (Mart.) Mart.	H	UB109322
Lythraceae	<i>Cuphea spermacoce</i> A.St.-Hil.	S	UB109338
Malpighiaceae	<i>Banisteriopsis anisandra</i> (A.Juss.) B.Gates	C	UB109174
	<i>Banisteriopsis argyrophylla</i> (A.Juss.) B.Gates	C	UB109455
	<i>Banisteriopsis stellaris</i> (Griseb.) B.Gates	C	UB109427

	<i>Diplopterys pubipetala</i> (A.Juss.) W.R.Anderson & C.C.Davis.	C	UB109233
Malvaceae	<i>Pavonia rosa-campestris</i> A.St.-Hil.	H	UB109213
	<i>Sida linifolia</i> Cav.	S	UB111995
Menispermaceae	<i>Cissampelos pareira</i> L.	C	UB110712
Myrsinaceae	<i>Cybianthus densiflorus</i> Miq.	S	UB109222
	<i>Cybianthus detergens</i> Mart.	S	UB109345
Myrtaceae	<i>Myrcia myrtilifolia</i> DC.	S	UB109269
Ochnaceae	<i>Ouratea floribunda</i> (A.St.-Hil.) Engl.	S	UB109570
	<i>Ouratea riedeliana</i> Engl.	S	UB109641
Orchidaceae	<i>Gomesa varicosa</i> (Lindl.) M.W.Chase & N.H.Williams	H	UB109414
Oxalidaceae	<i>Oxalis suborbiculata</i> Lourteig	S	UB109302
Passifloraceae	<i>Passiflora cerradensis</i> Sacco	C	UB110713
Poaceae	<i>Aristida pendula</i> Longhi-Wagner	H	UB109619
	<i>Axonopus siccus</i> (Nees) Kuhlms.	H	UB109620
	<i>Echinolaena inflexa</i> (Poir.) Chase	H	UB109435
	<i>Hiladaea pallens</i> (Sw.) C.Silva & R.P.Oliveira	H	UB109384
	<i>Megathyrsus maximus</i> (Jacq.) B.K.Simon & S.W.L.Jacobs *	H	UB111988
	<i>Melinis minutiflora</i> P. Beauv. *	H	UB109383
	<i>Mesosetum loliiforme</i> (Hochst. ex Steud.) Chase	H	UB111990
	<i>Panicum cervicatum</i> Chase	H	UB109463
	<i>Panicum sellowii</i> Nees	H	UB109506
	<i>Paspalum polyphyllum</i> Nees	H	UB109617
	<i>Schizachyrium tenerum</i> Nees	H	UB111991
	<i>Trachypogon</i> sp.1	H	UB111993
	<i>Trachypogon</i> sp.2	H	UB111994
	<i>Urochloa brizantha</i> (Hochst. ex A.Rich.) R.D.Webster *	H	UB111985
	<i>Urochloa decumbens</i> (Stapf) R.D.Webster *	H	UB109360
	<i>Urochloa</i> sp. *	H	UB111986
Polygalaceae	<i>Asemeia violacea</i> (Aubl.) J.F.B.Pastore & J.R.Abbott	H	UB109353
Rubiaceae	<i>Borreria poaya</i> (A.St.-Hil.) DC.	H	UB109470
	<i>Chomelia ribesioides</i> Benth. ex A.Gray	S	UB109330
	<i>Coccocypselum aureum</i> (Spreng.) Cham. & Schltld.	H	UB109494
	<i>Declieuxia fruticosa</i> (Willd. ex Roem. & Schult.) Kuntze	S	UB109478
	<i>Galium noxium</i> (A.St.-Hil.) Dempster	H	UB109361
	<i>Palicourea hoffmannseggiana</i> (Willd. ex Schult.) Borhidi	S	UB109489
	<i>Palicourea marcgravii</i> A.St.-Hil.	S	UB109380
	<i>Palicourea officinalis</i> Mart.	S	UB109375
	<i>Palicourea prunifolia</i> (Kunth) Borhidi	S	UB109368
	<i>Palicourea trichophora</i> (Müll. Arg.) Delprete & J.H.Kirkbr.	S	UB109387
	<i>Sabicea brasiliensis</i> Wernham	S	UB109430
Sapindaceae	<i>Serjania erecta</i> Radlk.	C	UB109471
	<i>Serjania lethalis</i> A.St.-Hil.	C	UB109256
	<i>Serjania ovalifolia</i> Radlk.	C	UB109207
Smilacaceae	<i>Smilax brasiliensis</i> Spreng.	C	UB109436
Turneraceae	<i>Turnera lamiifolia</i> Cambess.	H	UB109584
Verbenaceae	<i>Lippia rotundifolia</i> Cham.	S	UB109607
	<i>Stachytarpheta polyura</i> Schauer	H	UB109389

Comparing the diversity of species in the interior of the cerradão with the edge of the same fragment found a higher value in the edge environment (H' edge = 2.26 nats/ind.; H' interior = 1.50 nats/ind).

Phytosociology of the herbaceous-subshrub layer

Phytosociological sampling of the herbaceous-subshrub layer identified 18 families, including three that were indeterminate, and 29 species, including 24 that were on the edge and 11 in the interior plots. The herbaceous *Rhynchospora exaltata* Kunth and *Echinolaena inflexa* (Poir.), the climbers *Dioscorea dodecaneura* Vell, and the subshrubs *Myrcia myrtillifolia* DC., *Palicourea hoffmannseggiana* (Willd. ex Schult.) Borhidi, and *Palicourea trichophora* Müll. Arg were present in both environments. Four species found at the edge of the cerradão were naturalized exotic, accounting for 13.79% of the total sample. No invasive species were observed in the plots within the fragment (Table 2).

Table 2. Species identified by phytosociological sampling of the herbaceous-subshrub stratum of the cerradão in the “REBIO do Cerradão”, Federal District, Brazil. I: interior; E: edge; *: Exotic naturalized; •: Not collected.

Family	Species	I	E	Voucher
Apocynaceae	<i>Ditassa obcordata</i> Mart.	X		UB109650
Arecaceae	<i>Allagoptera campestris</i> (Mart.) Kuntze	X		UB109299
Asteraceae	<i>Achyrocline satureioides</i> (Lam.) DC.		X	UB109643
Cyperaceae	<i>Rhynchospora exaltata</i> Kunth	X	X	UB109143
Dioscoreaceae	<i>Dioscorea dodecaneura</i> Vell.	X	X	UB110707
Fabaceae	<i>Stylosanthes guianensis</i> (Aubl.) Sw.		X	UB109210
Lauraceae	<i>Cassytha filiformis</i> L.		X	UB109460
Malpighiaceae	<i>Banisteriopsis stellaris</i> (Griseb.) B.Gates	X		•
Malvaceae	<i>Sida linifolia</i> Cav.		X	UB111995
Menispermaceae	<i>Cissampelos pareira</i> L.		X	UB110712
Myrtaceae	<i>Myrcia myrtillifolia</i> DC.	X	X	UB109269
Oxalidaceae	<i>Oxalis suborbiculata</i> Lourteig		X	UB109302
Poaceae	<i>Axonopus siccus</i> (Nees) Kuhlman.		X	UB109620
	<i>Echinolaena inflexa</i> (Poir.) Chase	X	X	UB109435
	<i>Hildebrandia pallens</i> (Sw.) C.Silva & R.P.Oliveira	X		UB109384
	<i>Megathyrsus maximus</i> (Jacq.) B.K.Simon & S.W.L.Jacobs *		X	UB111988
	<i>Melinis minutiflora</i> P. Beauv. *		X	UB109383
	<i>Mesosetum loliforme</i> (Hochst. ex Steud.) Chase		X	UB111990
	<i>Schizachyrium tenerum</i> Nees		X	UB111991
	<i>Trachypogon</i> sp.1		X	UB111993
	<i>Trachypogon</i> sp.2		X	UB111994
	<i>Urochloa brizantha</i> (Hochst. ex A.Rich.) R.D. Webster *		X	UB111985
	<i>Urochloa</i> sp. *		X	UB111986
Rubiaceae	<i>Palicourea hoffmannseggiana</i> (Willd. ex Schult.) Borhidi	X	X	UB109489
	<i>Palicourea trichophora</i> (Müll.Arg.) Delprete & J.H.Kirkbr.	X	X	UB109387
Verbenaceae	<i>Stachytarpheta polyura</i> Schauer		X	UB109389
Indetermined	Indetermined 1		X	•

Indetermined 2	X	•
Indetermined 3	X	•

Greater coverage was observed in the border plots than in the interior. The results of the Kolmogorov-Smirnov test were significant ($D = 0.475$; $P = 0.000133$), with different distributions identified between the sampled environments. Similarly, the Sørensen index ($S_s = 0.343$) revealed that the environments were not floristically similar, since values less than 0.5 are considered indicative of low similarity (Kent, 2012).

Poaceae was the richest family in terms of species and constituted the largest number of sampled representatives (11 spp.). It presented phytosociological prominence, with species among the most frequent and dominant. In the interior plots, *E. inflexa* (“capim-flexinha”) and *R. exaltata* (“capim-estrela”) stood out in terms of frequency and dominance. On the border, *Melinis minutiflora* P. Beauv. and *Urochloa brizantha* (Hochst. ex A.Rich.) Webster - both exotic - were the most dominant species, although *E. inflexa* occurred in more than half of the plots sampled in this environment (Table 3).

Table 3. Phytosociological parameters of species present in samples of the herbaceous-subshrub stratum from the cerradão in the “REBIO do Cerradão”, Federal District, Brazil, in descending order of the total value of relative dominance. FA: absolute frequency (%); DoR: relative dominance (%). •: Absent.

Species	Edge		Interior		Total	
	FA	DoR	FA	DoR	FA	DoR
<i>Echinolaena inflexa</i>	52,5	11,38	67,5	51,53	60	62,91
<i>Rhynchospora exaltata</i>	27,5	7,77	80	39,56	53,75	47,33
<i>Urochloa brizantha</i>	37,5	21,23	•	•	18,75	21,23
<i>Melinis minutiflora</i>	42,5	19,19	•	•	21,25	19,19
<i>Myrcia myrtilifolia</i>	2,5	0,94	20	2,70	11,25	3,64
<i>Palicourea hoffmannseggiana</i>	2,5	0,47	7,5	2,99	5	3,46
<i>Palicourea trichophora</i>	2,5	1,40	5	1,61	3,75	3,01
<i>Ichnanthus bambusiflorus</i>	7,5	2,43	•	•	3,75	2,43
<i>Axonopus siccus</i>	7,5	2,11	•	•	3,75	2,11
<i>Cassytha filiformis</i>	5	2,11	•	•	2,5	2,11
<i>Megathyrsus maximus</i>	5	2,11	•	•	2,5	2,11
<i>Achyrocline satureioides</i>	2,5	1,87	•	•	1,25	1,87
Indetermined 1	2,5	1,87	•	•	1,25	1,87
Indetermined 2	5	1,08	•	•	1,25	1,08
<i>Trachypogon</i> sp.1	15	1,03	•	•	7,5	1,03
<i>Cissampelos pareira</i>	2,5	0,47	•	•	1,25	0,47
<i>Mesosetum loliiforme</i>	2,5	0,47	•	•	1,25	0,47
<i>Hildaea pallens</i>	•	•	2,5	0,44	1,25	0,44
<i>Trachypogon</i> sp.2	2,5	0,33	•	•	1,25	0,33
<i>Allagoptera campestris</i>	•	•	2,5	0,29	1,25	0,29
<i>Ditassa obcordata</i>	•	•	2,5	0,29	1,25	0,29

Indetermined 3	•	•	5	0,29	3,75	0,29
<i>Schizachyrium tenerum</i>	2,5	0,23	•	•	1,25	0,23
<i>Stachytarpheta polyura</i>	2,5	0,23	•	•	1,25	0,23
<i>Dioscorea dodecaneura</i>	2,5	0,05	5	0,15	1,25	0,19
<i>Banisteriopsis cf. stellaris</i>	•	•	2,5	0,15	1,25	0,15
<i>Sida linifolia</i>	2,5	0,09	•	•	1,25	0,09
<i>Stylosanthes guianensis</i>	2,5	0,09	•	•	2,5	0,09
<i>Oxalis suborbiculata</i>	2,5	0,05	•	•	2,5	0,05

The invasive grass *M. minutiflora* (“capim-gordura”) occurred in the portion of the cerradão edge closest to the adjacent dense savanna, where the vegetation is lower due to the presence of smaller species that favor the reach of light in the lower strata of vegetation. At the edge of the cerradão, *M. minutiflora* and *E. inflexa* do not seem to be excluded, since the species share 80% of the plots occupied by *M. minutiflora*. Conversely, *U. brizantha* (“braquiária”) was the only species able to stop its spread, since in the plots where it dominated, *M. minutiflora* was absent or presented low values of dominance.

R. exaltata, a common species in areas of cerradão (Mendonça *et al.*, 2008), was present in 80% of the plots in the interior and 27.5% of the plots on the edge of the cerradão, where it was represented by a large number of small individuals.

Dead biomass was only recorded on the edge, comprising up to 30% of the sampled biomass. *E. inflexa* and *M. minutiflora* accounted for 87.35% of the total mulching. In the interior of the fragment, where the environment is widely covered by the treetops, coverage of the herbaceous-subshrub stratum hardly exceeded 50%, while coverage of half the plots allocated at the edge exceeded 50%. Despite this, there was a greater number of plots on the edge (four plots) with less than 1% coverage compared to the interior (three plots) (Table 4).

Table 4. Number of plots by coverage class sampled in the cerradão in the “REBIO do Cerradão”, Federal District, Brazil.

Braun-Blanquet coverage classes	Nº of plots	
	Edge	Interior
< 1%	4	3
1 - 5%	1	9
6 - 25%	7	17
26 - 50%	8	10
51 - 75%	6	0
76 - 100%	14	1
Total	40	40

DISCUSSION

Richness and diversity

Studies sampling the herbaceous stratum in cerradão are rare, therefore, limited data are available for comparison. However, the cerradão studied at “REBIO do Cerradão” presented high floristic richness, when compared to other cerradão areas (109 spp.). In the state of São Paulo, Batalha and Mantovani (2001) found 70 herbaceous-subshrub species, of which almost 3% were invasive. Rubiaceae (nine species), Asteraceae, and Poaceae (six species each) presented greater richness, and 18.6% of the total families were represented by only one species. In South Mato Grosso, Assunção *et al.* (2011) cataloged

59 species, 41.2% of families were represented by only one species, and 10.17% of subspontaneous species. Fabaceae was the richest family (15 spp.), followed by Poaceae (12 spp.), and Asteraceae (seven spp.). In “REBIO do Cerradão” the proportion of families represented by a single species (60.52%) is higher than reported in other areas of cerradão. Fabaceae, Poaceae, and Asteraceae are also among the richest families.

In the literature, these families are identified as the richest families in the Brazilian savanna and are best represented in the herbaceous communities of cerradão (Filgueiras, 2002; Mendonça *et al.*, 2008). Assunção *et al.* (2011) found that species belonging to the Asteraceae family were more common at the edges and in the clearings present inside the fragment. Although the family was represented by 14 species in the floristic survey of this cerradão, only one species (*Achyrocline satureioides* [Lam.] DC.) was sampled in a phytosociological study. Despite being native to Brazil, *A. satureioides* is common in open and altered environments, such as wastelands and roadsides (Moreira & Bragança, 2011).

Asteraceae is one of the main families of invasive plants, and is particularly common in open formations in Brazil, mainly in the Brazilian savanna (Souza & Lorenzi, 2019). According to Coutinho (1978), Asteraceae and Poaceae, together with Cyperaceae, include genera and species with a large number of heliophile representatives. In addition, that author noted the herbaceous-subshrub layer is predominantly composed of heliophilous perennial species and is poorly represented in the cerradão due to intolerance to shading. In the present study, the richness of this stratum was probably due to the presence of heliophilous species in the border environment and in the clearings within the fragment, where the incidence of light was greater.

Phytosociology of the herbaceous-subshrub layer

In the cerradão, there was greater coverage in the border environment. *E. inflexa* and *M. minutiflora* were among the most important herbaceous species in all cerradão areas studied by Felfili *et al.* (1994) in Chapada Pratinha, in the Brazilian Central Plateau. Pivello *et al.* (1999) recorded the highest frequency, dominance, and density values for *E. inflexa* among native herbs observed in a savanna reserve in the state of São Paulo.

According to Hoffmann and Haridasan (2008), *M. minutiflora* prefers shaded areas and tends to form a dense cover of approximately 1 m in height, which inhibits the survival of tree seedlings. However, at the edge of REBIO's cerradão, *E. inflexa* was present in 80% of the plots occupied by *M. minutiflora*; this was also observed by Pivello *et al.* (1999).

“Capim-gordura” is not efficient at exploring extensive areas of soil due to the small size of its roots; however, some variables, such as increased nutrient mineralization resulting from burning and/or increased organic matter, can improve nutrient availability on the soil surface. In these places, due to its capacity for vegetative expansion, the invasive species is able to replace the native species, thus reducing local biodiversity (Martins *et al.*, 2004). In the present study, no seedlings or subshrubs were found in the plots where *M. minutiflora* was present at a high density.

In addition to *E. inflexa*, *R. exaltata* stood out in the interior plots. Filgueiras (2002) and Alves *et al.* (2009) reported that the family Cyperaceae, more specifically *R. exaltata*, is characteristic of open areas, and tends to appear in high densities in forest edges and clearings, in humid areas of forest (edge and interior), and between outcrops of quartzitic fields (Longhi-Wagner & Araújo, 2014). In a dystrophic seasonal forest, Meira-Neto *et al.* (2005) found a positive correlation between *R. exaltata* and low coverage, high brightness, and soils with high exchangeable aluminum content and low exchangeable Ca, Mg and K levels; these edaphic characteristics are also observed in “REBIO do Cerradão.”

Pivello *et al.* (1999) noted that *R. exaltata* acts as a background herb on which other species occupy smaller areas, except in the presence of “braquiária” (*Urochloa decumbens*

(Stapf) R. D. Webster). Of the 15 plots (37.5%) containing *U. brizantha*, which were all on the edge, *R. exaltata* was only present in one, with 3% coverage of the plot. Non-native herbs tend to disperse easily due to anemocory, have higher rates of photosynthesis, and are more efficient at using nutrients and nitrogen compared to native species (Eller & Oliveira, 2007; Silva & Haridasan, 2017).

Among the subshrub species sampled, *Achyrocline satureioides*, *Sida linifolia* Cav., *Oxalis suborbiculata* Lourteig, *Palicourea trichophora*, and *P. hoffmannseggiana* are common to anthropized areas and/or forest edges (Mendonça *et al.*, 2008), and the latter is one of the most frequent species in the cerrado areas of Chapada Pratinha (Felfili *et al.*, 1994). Within the studied fragment, the most frequent subshrub, present in eight plots, was *Myrcia myrtillofolia*, which is a native species commonly found in areas of woodland savanna (i.e., cerrado *sensu stricto*), rupestrian savanna, and grassland (Proença *et al.*, 2001; Mendonça *et al.*, 2008; Lucas *et al.*, 2016).

The presence of exotic species in the border plots only, totaling more than 10% of the total species observed in this environment and contributing 53% of its live cover, seems to corroborate the statement that the fragmentation of an area facilitates the invasion process, hindering, or even preventing the development of naturally occurring species. Felfili *et al.* (1994) found that the cerrado areas presented the highest number of invasive species, highlighting the vulnerability of phytophysiology. The high number of invasive species was also responsible for the greater richness and diversity of the border compared to the interior of the fragment, which reflects the low floristic similarity between the sampled environments.

The lower coverage in the interior plots was expected, since herbaceous species have a preference for more open environments where light falls more abundantly. This also explains why 10% of the plots on the edge have less than 1% coverage, because the portion of the edge where these plots were allocated was close to the water reservoir; therefore, it was more recent and shaded than the rest of the strip sampled. In this environment, the herbaceous layer is null or reduced and the tree species present form a canopy, which provides constant shading. As this occurs inside the fragment, it appears to hinder the establishment of herbaceous species that require high luminosity.

Research on the herbaceous-subshrub stratum in the cerrado is important for expanding our knowledge on this specific vegetation and for conservation purposes. It is a stratum that is floristically little known in phytophysiology, as observed by the limited number of studies with this objective. Because they are smaller, they face greater competition for light, nutrients, and water, and possess structural and physiological adaptations associated with the environment in which they live. Plants in the lower stratum become sensitive to changes in this environment, and act as indicators of environmental quality, providing useful data on the conservation status of forest communities (Müller & Waechter, 2001; Miyamura *et al.*, 2019).

For the Brazilian savanna, little is known about vegetation dynamics in an edge environment since most studies have been performed in areas of tropical and temperate forests (Turner, 1996; Nascimento & Laurance, 2006; Laurance & Vasconcelos, 2009; Smith *et al.*, 2018). Based on the limited studies carried out in the Brazilian savanna, there is a wide variety of responses from the organisms, and the edge may or may not influence the structure and composition of the fragmented area (Lima-Ribeiro, 2008; Sampaio & Scariot, 2011; Arruda & Eisenlohr, 2016).

CONCLUSIONS

Considering the difference in the coverage and composition of the flora of the herbaceous-subshrub layer between the two distances evaluated, since inside the fragment

the vegetation remains closer to the original formation with species typical of phytophysiology, and that the native species of the biome that occur on the edge are species that tend to appear in anthropized areas or edges of forests, we conclude that the cerrado of REBIOS presents fragmentation effects. However, a future study in this fragment comparing the data obtained here, may provide more definitive answers.

REFERENCES

- ALVES, M.; ARAÚJO, A.C.; PRATA, A.P.; VITTA, F.A.; HEFLER, S.M.; TREVISAN, R.; GIOL, A.S.B.; MARTINS, S. & THOMAS, W. 2009. Diversity of Cyperaceae in Brazil. *Rodriguésia* 60(4): 771-782. Doi: 10.1590/2175-7860200960405.
- ANGIOSPERM PHYLOGENY GROUP IV. 2020. An update of the Angiosperm Phylogeny Group classification for the orders and families of flowering plants: APG IV. *Botanical Journal of the Linnean Society*. doi: 10.1111/boj.12385.
- AQUINO, F.G. & MIRANDA, G.H.B. 2008. Consequências ambientais da fragmentação de habitats no Cerrado. In: S.M. Sano, S.P. Almeida & J.F. Ribeiro, (eds). *Cerrado: Ecologia e Flora*. v.1. Embrapa Cerrados, Brasília. pp. 383-398.
- ARRUDA, D.M. & EISENLOHR, P.V. 2016. Analyzing the edge effects in a Brazilian seasonally dry tropical forest. *Brazilian Journal of Biology* 76: 169-175. Doi: 10.1590/1519-6984.16014.
- ASSUNÇÃO, V.A.; GUGLIERI-CAPORAL, A. & SARTORI, A.L.B. 2011. Florística do estrato herbáceo de um remanescente de cerrado em Campo Grande, Mato Grosso do Sul, Brasil. *Hoehnea* 38(2): 281-288.
- BATALHA, M.A. & MANTOVANI, W. 2001. Floristic composition of the Cerrado in the Pé-de-Gigante Reserve (Santa Rita do Passa Quatro, Southeastern, Brazil). *Acta Botanica Brasilica* 15: 289-304.
- BEUCHLE, R.; GRECCHI, R.C.; SHIMABUKURO, Y.E.; SELIGER, R.; EVA, H.D.; SANO, E. & ACHARD, F. 2015. Land cover changes in the Brazilian Cerrado and Caatinga biomes from 1990 to 2010 based on a systematic remote sensing sampling approach. *Applied Geography* 58: 116-127.
- BRASIL. 2000. Lei Federal n. 9.985, de 18 de julho de 2000. *Sistema Nacional de Unidade de Conservação da Natureza – SNUC*. Presidência da República / Casa Civil. Available at http://www.planalto.gov.br/ccivil_03/leis/l9985.htm.
- BRAZIL FLORA GROUP. 2021. *Brazilian Flora 2020 Project / Projeto Flora do Brasil 2020*. v393.274. Instituto de Pesquisas Jardim Botânico do Rio de Janeiro. Dataset/Checklist. doi:10.15468/1mtkaw.
- CARDOSO, M.R.D.; MARCUZZO, F.F.N. & BARROS, J.R. 2014. Classificação climática de Köppen-Geiger para o Estado de Goiás e o Distrito Federal. *Acta Geografica* 8(16): 40-55.
- CLIMATE-DATA. 2020. *Brasília Clima (Brasil)*. Available at <https://pt.climate-data.org/americas-do-sul/brasil/distrito-federal/brasilia-852>. Accessed in 09 October 2020.
- CONSTANTINO, R.; BRITTEZ, R.M.; CERQUEIRA, R.; ESPINDOLA, E.L.G.; GRELE, C.E.V.; LOPES, A.T.L.; NASCIMENTO, M.T.; ROCHA, O.; RODRIGUES, A.A.F.; SCARIOT, A.; SEVILHA, A.C. & TIEPOLO, G. 2005. Causas naturais. In: D.M. Rambaldi & D.A.S. Oliveira, (orgs). *Fragmentação de ecossistemas: causas, efeitos sobre a biodiversidade e recomendações de políticas públicas*. 2ª ed. Ministério do Meio Ambiente / Secretaria de Biodiversidade e Florestas, Brasília. pp. 44-63.
- COUTINHO, L.M. 1978. O conceito de Cerrado. *Revista Brasileira de Botânica* 7: 17-23.

- COUTINHO, L.M. 2006. O conceito de bioma. *Acta Botanica Brasílica* 20(1): 13-23.
- CHRISTIANINI A.V. & OLIVEIRA, P.S. 2013. Edge effects decrease ant-derived benefits to seedlings in a neotropical savanna. *Arthropod-Plant Interactions* 7:191–199. Doi: 10.1007/s11829-012-9229-9.
- DIAMOND, J.M. 1976. Island Biogeography and Conservation: strategy and limitations. *Science* 193: 1027-1029.
- DISTRITO FEDERAL. 1998. Decreto n. 19.213, de 06 de maio de 1998. *Diário Oficial do Distrito Federal*, 07 May 1998, sec. 1, pp. 1-2.
- DISTRITO FEDERAL. 2010. Decreto n. 31.757, de 02 de junho de 2010. *Diário Oficial do Distrito Federal*, 07 June 2010, sec. 1, p. 20.
- DODONOV, P.; HARPER, K.A. & SILVA-MATOS, D.M. 2013. The role of edge contrast and forest structure in edge influence: vegetation and microclimate at edges in the Brazilian cerrado. *Plant Ecology* 214: 1345-1359.
- DODONOV, P.; BRAGA, A.L.; HARPER, K.A. & SILVA-MATOS, D.M. 2016. Edge influence on plant litter biomass in forest and savanna in the Brazilian cerrado. *Austral Ecology* 42: 187-197.
- ELLER, C.B. & OLIVEIRA, R.S. 2017. Effects of nitrogen availability on the competitive interactions between an invasive and a native grass from Brazilian cerrado. *Plant and Soil* 410: 63-72. Doi: 10.1007/s11104-016-2984-0.
- FELFILI, J.M.; HARIDASAN, M.; MENDONÇA, R.C.; FILGUEIRAS, T.S.; SILVA JUNIOR, M.C. & REZENDE, A.V. 1994. Projeto Biogeografia do Bioma Cerrado: vegetação e solos. *Caderno de Geociências* 12: 75-166.
- FILGUEIRAS, T.S. 2002. Herbaceous plant communities. In: P.S. Oliveira & R.J. Marquis, (eds). *The Cerrados of Brazil: ecology and natural history of a neotropical savanna*. Columbia University Press, New York. pp. 121-139.
- HAMMER, O.; HARPER, D.A.T. & RYAN, P.D. 2008. PAST: Paleontological Statistics. ver. 1.81. Available at <http://folk.uio.no/ohammer/past>.
- HOFFMANN, W.A. & HARIDASAN, M. 2008. The invasive grass, *Melinis minutiflora*, inhibits tree regeneration in a Neotropical savanna. *Austral Ecology* 33: 29-36.
- INSTITUTO BRASÍLIA AMBIENTAL. 2014. Mapa Ambiental do Distrito Federal – ano 2014, SEMARH-DF/GDF. Available at http://www.ibram.df.gov.br/wp-content/uploads/2018/02/Mapa_Ambiental_20141.pdf. Accessed in 01 October 2020.
- KENT, M. 2012. Vegetation description and data analysis: a practical approach. Willey-Blackwell, London.
- KOVACH, W.L. 2007. *MVSP - A MultiVariate Statistical Package for Windows*. ver. 3.1. Kovach Computing Services.
- LAURANCE, W.F. & VASCONCELOS, H.L. 2009. Consequências ecológicas da fragmentação florestal na Amazônia. *Oecologia Brasiliensis* 13: 434-451. Doi: 10.4257/oeco.2009.1303.030.
- LIMA-RIBEIRO, M.S. 2008. Efeitos de borda sobre a vegetação e estruturação populacional em fragmentos de cerrado no Sudoeste Goiano, Brasil. *Acta Botanica Brasílica* 22(2): 535-545.
- LONGHI-WAGNER, H.M. & ARAÚJO, A.C. 2014. Flora fanerogâmica da Serra do Ouro Branco, Minas Gerais, Brasil: Cyperaceae. *Rodriguésia* 65(2): 369-404.
- LUCAS, E.; WILSON, C.E.; LIMA, D.F.; SOBRAL, M. & MATSUMOTO, K. 2016. A conspectus of *Myrcia* sect. *Aulomyrcia* (Myrtaceae). *Annals of the Missouri Botanical Garden* 101(4): 648-698.
- MACARTHUR, R.H. & WILSON, E.O. 2016. *The Theory of Island Biogeography*. Princeton University Press, New Jersey.

- MARTINS, C.R.; LEITE, L.L. & HARIDASAN, M. 2004. Capim-gordura (*Melinis minutiflora* P.Beauv.), uma gramínea exótica que compromete a recuperação de áreas degradadas em Unidades de Conservação. *Revista Árvore* 28(5): 739-747.
- MEIRA-NETO, J.A.A. & MARTINS, F.R. 2000. Composição florística do estrato herbáceo-arbustivo de uma floresta estacional semidecidual em Viçosa-MG. *Revista Árvore* 24(4): 407-416.
- MEIRA-NETO, J.A.A.; MARTINS, F.R. & SOUZA, A.L. 2005. Influência da cobertura e do solo na composição florística do sub-bosque em uma floresta estacional semidecidual em Viçosa, MG, Brasil. *Acta Botanica Brasilica* 19(3): 473-486.
- MENDONÇA, R.C.; FELFILI, J.M.; WALTER, B.M.T.; SILVA JUNIOR, M.C.; REZENDE, A.B.; FILGUEIRAS, T.S.; NOGUEIRA, P.E. & FAGG, C.W. 2008. Flora vascular do bioma Cerrado: *checklist* com 12.356 espécies. In: S.M. Sano, S.P. Almeida & J.F. Ribeiro, (eds). *Cerrado: Ecologia e Flora*. v.2. Embrapa Cerrados, Brasília. pp. 423-1279.
- MENDONÇA, A.H.; RUSSO, C.; MELO, A.C.G.; DURIGAN, G. 2015. Edge effects in savanna fragments: a case study in the Cerrado. *Plant Ecology & Diversity* 8: 493-503. Doi: 10.1080/17550874.2015.1014068.
- MIGUEL, E.P.; REZENDE, A.V.; PEREIRA, R.S.; AZEVEDO, G.B.; MOTA, F.C.M.; SOUZA, A.N. & JOAQUIM, M.S. 2017. Modeling and prediction of volume and aerial biomass of the tree vegetation in a cerradão area of central Brazil. *Interciencia* 42(1): 21-27.
- MIYAMURA, F.Z.; MANFRA, R.; FRANCO, G.A.D.C.; ESTEVES, R.; SOUZA, S.C.P.M. & IVANAUSKAS, N.M. 2019. Influência de espécies exóticas invasoras na regeneração natural de um fragmento florestal urbano. *Scientia Plena* 15: 082401. Doi: 10.14808/sci.plena.2019.08240.
- MOREIRA, H.J.C. & BRAGANÇA, H.B.N. 2011. *Manual de identificação de plantas infestantes: hortifrúti*. FMC Agricultural Products, São Paulo.
- MÜLLER, S.C. & WAECHTER, J.L. 2001. Estrutura sinusal dos componentes herbáceo e arbustivo de uma floresta costeira subtropical. *Revista Brasileira de Botânica* 24(4): 395-406.
- MÜELLER-DOMBOIS, D. & ELLENBERG, H. 2002. *Aims and methods of vegetation ecology*. The Blackburn Press, New Jersey.
- NASCIMENTO, H.E.M. & LAURANCE, W.F. 2006. Efeitos de área e de borda sobre a estrutura florestal em fragmentos de floresta de terra-firme após 13-17 anos de isolamento. *Acta Amazonica* 36(2): 183-192.
- PEREIRA, F.C.; GUILHERME, F.A.G. & MARIMON, B.S. 2021. Edge effects on successional dynamics of forest fragments in the Brazilian Cerrado. *Floresta e Ambiente* 28(2): e20200063.
- PIELOU, E.C. 1966. Species diversity and pattern diversity in the study of ecological succession. *Journal of Theoretical Biology* 10(2): 370-383.
- PIVELLO, V.R.; CARVALHO, V.M.C.; LOPES, P.F.; PECCININI, A.A. & ROSSO, S. 1999. Abundance and distribution of native and alien grasses in a "Cerrado" (Brazilian Savanna) biological reserve. *Biotropica* 31(1): 71-82.
- PRIMACK, R.B. & RODRIGUES, E. 2001. *Biologia da Conservação*. Planta, Londrina.
- PROENÇA, C.E.B.; MUNHOZ, C.B.R.; JORGE, C.L. & NÓBREGA, M.G.G. 2001. Listagem e nível de proteção das espécies de fanerógamas do Distrito Federal, Brasil. In: T.B. Cavalcanti & A.E. Ramos, (eds). *Flora do Distrito Federal I*. Embrapa Recursos Genéticos e Biotecnologia, Brasília. pp. 87-359.
- RIBEIRO, J.F. & WALTER, B.M.T. 2008. As principais fitofisionomias do bioma Cerrado. In: S.M. Sano, S.P. Almeida & J.F. Ribeiro, (eds). *Cerrado: Ecologia e Flora*. v.1. Embrapa Cerrados, Brasília. pp. 151-212.

- SAMPAIO, A.B. & SCARIOT, A. 2011. Efeito de borda na diversidade, composição e estrutura da comunidade arbórea em uma Floresta Estacional Decidual no Brasil Central. *Revista Árvore* 35: 1121-1134. Doi: 10.1590/S0100-67622011000600018.
- SANO, E.E.; ROSA, R.; BRITO, J.L.S. & FERREIRA, L.G. 2010. Land cover mapping of the tropical savana region in Brazil. *Environmental Monitoring and Assessment* 166(1-4): 113-124.
- SILVA, J.S.O. & HARIDASAN, M. 2007. Acúmulo de biomassa aérea e concentração de nutrientes em *Melinis minutiflora* P. Beauv. e gramíneas nativas do cerrado. *Revista Brasileira de Botânica* 30(2): 337-344.
- SILVA, J.S. & FELFILI, J.M. 2012. Floristic composition of a conservation area in the Federal District of Brazil. *Brazilian Journal of Botany* 35(4): 385-395.
- SMITH, I.A.; HUTYRA, L.R.; REINMANN, A.B.; MARRS, J.K. & THOMPSON, J.R. 2018. Piecing together the fragments: elucidating edge effects on forest carbon dynamics. *Frontiers in Ecology and the Environment* 16(4): 213-221.
- SOUZA, V.C. & LORENZI, H. 2019. *Botânica Sistemática: guia ilustrado para identificação das famílias de Fanerógamas nativas e exóticas no Brasil, baseado em APG IV*. Instituto Plantarum, São Paulo.
- STRASSBURG, B.B.N.; BROOKS, T.; FELTRAN-BARBIERI, R.; IRIBARREM, A.; CROUZEILLES, R.; LOYOLA, R.; LATAWIEC, A.E.; OLIVEIRA FILHO, F.J.B.; SCARAMUZZA, C.A.M.; SCARANO, F.R.; SOARES-FILHO, B. & BALMFORD, A. 2017. Moment of truth for the Cerrado hotspot. *Nature Ecology & Evolution*. doi:10.1038/s41559-017-0099.
- TALLAMY, D.W.; NARANGO, D.L.; MITCHELL, A.B. 2020. Do non-native plants contribute to insect declines? *Ecological Entomology*. doi:10.1111/een.12973.
- TURNER, I.M. 1996. Species loss in fragments of tropical rain forest: a review of the evidence. *Journal of Applied Ecology* 33(2): 200-209.
- URBANIAK, G.C. & PLOUS, S. 2007. *Research Randomizer*. ver. 3.0. Available at <http://www.randomizer.org/>. Accessed in 01 October 2008.
- VOGADO, N.O.; CAMARGO, M.G.G.; LOCOSSELLI, G.M. & MORELLATO, L.P.C. 2016. Edge effects on the phenology of the guamirim, *Myrcia guianensis* (Myrtaceae), a cerrado tree, Brazil. *Tropical Conservation Science* 9(1): 291-312.
- ZAR, J.H. 2010. *Biostatistical analysis*. Prentice Hall, New Jersey.

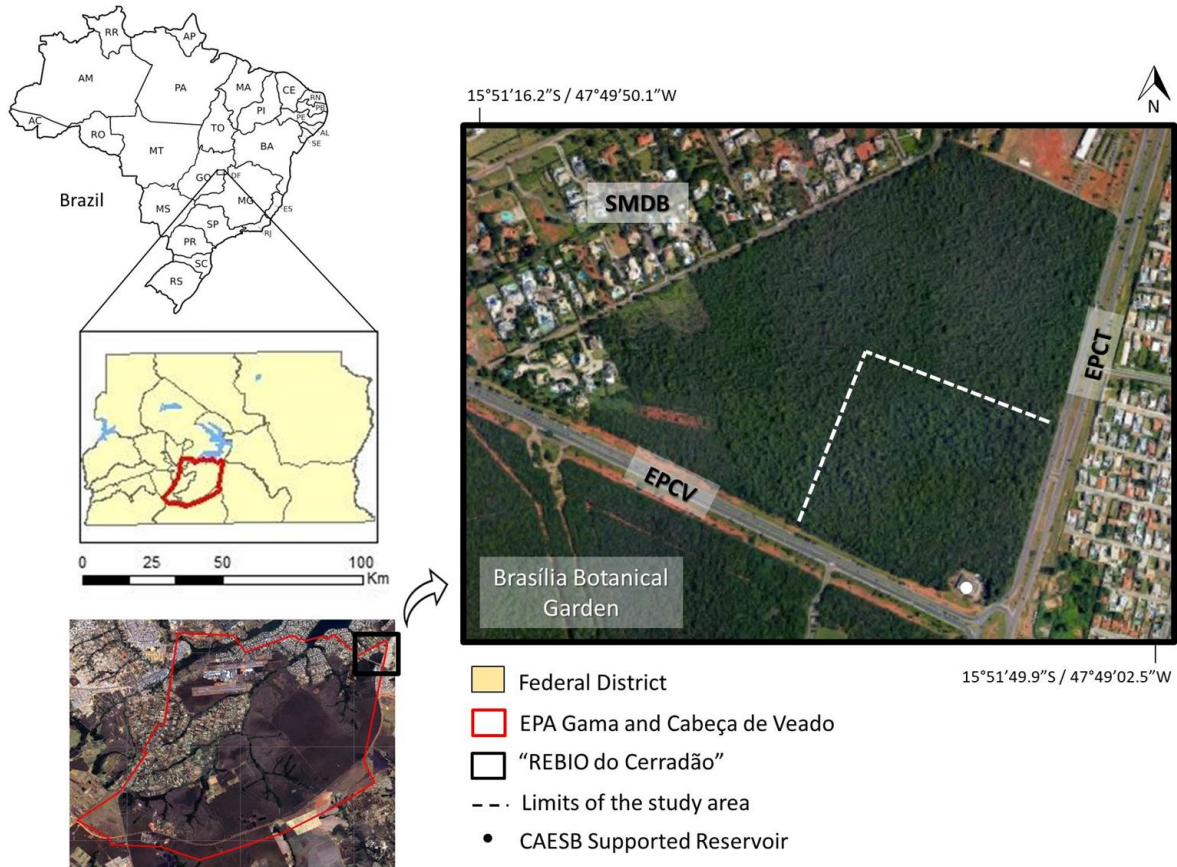


Figure 1. Study area location. Source: Modified from Instituto Brasília Ambiental (2014) and Google Earth (Year of the image: 2020).

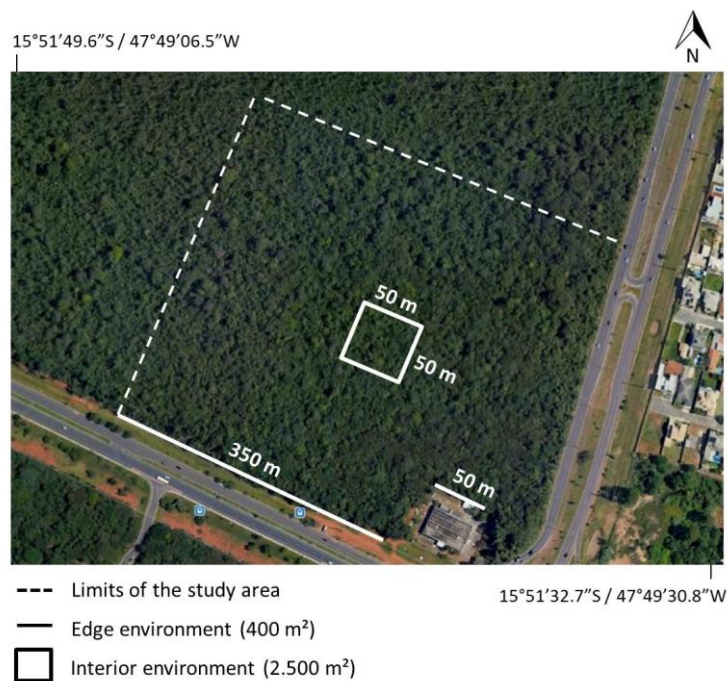


Figure 2. Edge and interior environments sampled at "REBIO do Cerradão", Federal District, Brazil, for a phytosociological study of the herbaceous-subshrub stratum. Source: Modified from Google Earth (Year of the image: 2020).